A review of the impacts of degradation threats on soil properties in the UK

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Abstract

National governments are becoming increasingly aware of the importance of their soil resources and are shaping strategies accordingly. Implicit in any such strategy is that degradation threats and their potential effect on important soil properties and functions are defined and understood. In this paper, we aimed to review the principal degradation threats on important soil properties in the UK, seeking quantitative data where possible. Soil erosion results in the removal of important topsoil and, with it, nutrients, C and porosity. A decline in soil organic matter principally affects soil biological and microbiological properties, but also impacts on soil physical properties because of the link with soil structure. Soil contamination affects soil chemical properties, affecting nutrient availability and degrading microbial properties, whilst soil compaction degrades the soil pore network. Soil sealing removes the link between the soil and most of the ‘spheres’, significantly affecting hydrological and microbial functions, and soils on re-developed brownfield sites are typically degraded in most soil properties. Having synthesized the literature on the impact on soil properties, we discuss potential subsequent impacts on the important soil functions, including food and fibre production, storage of water and C, support for biodiversity, and protection of cultural and archaeological heritage. Looking forward, we suggest a twin approach of field-based monitoring supported by controlled laboratory experimentation to improve our mechanistic understanding of soils. This would enable us to better predict future impacts of degradation processes, including climate change, on soil properties and functions so that we may manage soil resources sustainably.

Keywords: Soil erosion, soil organic matter, soil contamination, soil compaction, soil functions, climate change

Introduction

National governments are becoming increasingly aware of the importance of their soil resources and have begun to shape policies to recognize this. In the UK, for example, there have been soil strategies and policy documents in England (HMG, 2009b, 2011b), Scotland (Dobbie et al., 2011) and Wales (WAG, 2006). All have a similar central aim, namely the ambition to manage soils sustainably and tackle potential degradation threats so that their ability to provide essential services is protected or enhanced. Soil degradation in England has been estimated to have cost the economy at least £150M–£250M each year (HMG, 2011b).

Implicit in any such strategy is that the principal degradation threats and their effect on important soil properties are defined and understood. If this can be achieved then, by extension, we may better understand the likely impact on soil functions – the essential services that we rely on soils to perform – and we can begin to define policies...
to protect them. Evidence-gathering is an initial stage in the process of developing and refining policies (HMG, 2011a).

In this paper, we aimed to identify and review the principal degradation threats on important soil properties in the UK, seeking quantitative data where possible. We summarize the main findings and discuss the likely subsequent impacts on soil functions and also the potential impacts of predicted changes in climate in the UK.

Methodology

Degradation threats

In examining which degradation threats were important in the context of the UK, we consulted the report of the European Environmental Assessment of Soil for Monitoring (ENVASSO) project (Huber et al., 2008), which aimed to establish a soil monitoring system in support of a proposed European Union-wide Soil Framework Directive (EC, 2006). In the report of Huber et al. (2008), nine key threats to soil were identified: soil erosion, decline in soil organic matter (SOM), soil contamination, soil sealing, soil compaction, decline in soil biodiversity, soil salinization, landslides and desertification. In addition, the report identified three further ‘cross-cutting issues’ that exist as threats to soil: climate change, land use change and brownfield development. Similar threats to soils have been identified in the UK previously (e.g. SG, 2009; HMG, 2011b).

From this list, we focused on soil erosion, decline in SOM, soil contamination, soil compaction, soil sealing and brownfield development, as we believe these to be of the greatest relevance to soils in the UK. It is our view that the decline in soil biodiversity does not exist as a specific independent threat to soil, but that instead it arises as a result of some of the other threats identified. Similarly, land-use change imparts some of the other threats identified on the soil. Therefore, these two threats were covered in the context of the other threats. We review soil salinization and landslides only briefly, as we consider these to be less important, and extremely localized in the UK. Desertification is not a threat currently to UK soils and is not reviewed. Finally, we discuss climate change separately in terms of what the likely impacts might be on degradation threats and soil properties. The degradation threats considered and brief definitions are given in Table 1.

<p>| Table 1 The eight degradation threats identified as important for soils in the UK. The descriptions are those of Huber et al. (2008), unless stated otherwise |</p>
<table>
<thead>
<tr>
<th>Degradation threat</th>
<th>Brief description</th>
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<tbody>
<tr>
<td>Soil erosion</td>
<td>The accelerated loss of soil as a result of anthropogenic activity in excess of accepted rates of natural soil formation</td>
</tr>
<tr>
<td>Decline in soil organic matter</td>
<td>The negative imbalance between the build-up of soil organic matter and rates of decomposition, leading to an overall decline in soil organic matter content and/or quality</td>
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<tr>
<td>Soil contamination</td>
<td>The presence of a substance or agent in the soil as a result of human activity emitted from moving sources, from sources with a large area or from many sources (diffuse) (ISO, 2006), or where intensive industrial activities, inadequate waste disposal, mining, military activities or accidents introduce excessive amounts of contaminants (local)</td>
</tr>
<tr>
<td>Soil compaction</td>
<td>The densification and distortion of soil by which total and air-filled porosity are reduced</td>
</tr>
<tr>
<td>Soil sealing</td>
<td>The destruction or covering of soil by buildings, constructions and layers or other bodies of artificial material which may be very slowly permeable to water (Burghardt et al., 2004)</td>
</tr>
<tr>
<td>Brownfield development</td>
<td>The further development of developed land previously used for commercial or industrial purposes, including restoration of green afters-uses</td>
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<tr>
<td>Soil salinization</td>
<td>The excessive increase of water-soluble salts in soil through natural processes and human interventions</td>
</tr>
<tr>
<td>Landslides</td>
<td>The movement of a mass of rock, debris, artificial fill or soil down a slope under the force of gravity (Cruden &amp; Varnes, 1996)</td>
</tr>
<tr>
<td>Climate change</td>
<td>The large-scale, long-term shift in the planet’s weather patterns or average temperatures (Met Office, 2015)</td>
</tr>
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Literture review

A literature review of peer-reviewed publications was conducted using the web-based ISI Web of Knowledge™ search engine (Thomson Reuter, New York, USA), supplemented by other relevant published evidence (e.g. reports). We took each degradation threat as a search term in turn with ‘soil’ and ‘UK’ and sought data, predominantly from the 2000 to 2015 period, on the impacts on important soil physical, biological and chemical properties in the UK. In particular, we focused on general properties likely to be of direct relevance for the ability of the soil to carry out its principal functions, including food and fibre production, water and air filtration, and biodiversity support. We synthesized the most relevant information in our review below.

The impacts of degradation threats on soil properties in the UK

Soil erosion. Although soil erosion can be natural, current concerns relate to accelerated erosion, where the rate has
Degradation threats to soil properties in the UK

Decline in soil organic matter. In common with other temperate-zone countries, there has been a decline in SOM in some soils in the UK over a long period, predominantly as a result of changes in land use. Historically land has changed from native woodland to grazing land, and then on to annual arable systems. Since the 1940s, the area covered by permanent grassland has reduced and in arable systems, there has been a reduction in the use of manures and an increase in cultivation depth (Bellamy et al., 2008), a decrease in friability (Watts & Dexter, 1998) and a reduction in porosity (Riley et al., 1997; Gregory et al., 2009). This can lead to an undesirable domination of coarser clods (Riley et al., 2008), a decrease in friability (Watts & Dexter, 1998) and a reduction in porosity (Riley et al., 2008). SOM decline can reduce soil water retention (Kibblewhite et al., 2007; Johnston et al., 2009) by up to 10% for a difference in SOM content from 7% to 3% (Gregory et al., 2009). Arable soils of low SOM content are susceptible to slumping when wetted due to aggregate instability, leading to less infiltration and greater surface runoff (Whitmore et al., 2004).

The quantity and quality of SOM, as the primary food source, largely controls the soil microbial biomass (Fierer et al., 2009) and biodiversity (Orwin et al., 2006; Wardle et al., 2006). Loss of SOM can reduce the exchange of important nutrients such as N, P and S (Johnston et al., 2009). SOM loss from 5% to 2% over a 60-yr period at Rothamsted resulted in a 90% decrease in microbial biomass, but no significant effect on microbial diversity (Hirsch et al., 2009) or substrate utilization (Wu et al., 2012). Smaller fungal biomass (Gregory et al., 2009) and fungal-to-bacterial biomass ratios (Bardgett et al., 1996, 2007) have been found in soils of low SOM content compared with undisturbed and botanically rich grassland soils in the UK.

Soil organic matter also plays a key role in the ability of soils to buffer the effects of potentially toxic substances, in part by chelation and adsorption (Chander & Brookes, 1993; Hund-Rinke & Kördel, 2003; Griffiths et al., 2005, 2008; Kuan et al., 2007). Loss of SOM can, hence, release toxic elements (ROTAP, 2009).

Soil contamination. Soil contamination may result from point or diffuse sources. Impacts of the latter are hard to predict as it can be affected by factors such as the weather, and soils far away from the source may be affected. Contamination mainly affects soil biological and chemical properties, although some contaminants (e.g. salts) may destabilize soil structure and affect soil physical properties.
Contamination that decreases soil pH, such as N deposition (NEGTAP, 2001; ROTAP, 2009), reduces nutrient availability, even at moderate levels (pH 5–7), and increases the risk of nutrient leaching (Pearson & Stewart, 1993; Jeffries & Maron, 1997; Degryse et al., 2007). Soil pH is one of the main determinants of soil biodiversity and functioning (Fierer & Jackson, 2006; Smolders et al., 2009; Griffiths et al., 2011). Some keystone soil taxa including Lumbricidae (earthworms), Collembola (springtails) and N-fixing bacteria are particularly sensitive to metals (Emmett et al., 2010). Contamination can affect the health of soil invertebrates in general (Spurgeon & Hopkin, 1996). Addition of Cu to UK soils (equivalent to 500 mg/kg) decreased both microbial respiration rates by up to 80% initially, and microbial biomass itself (Griffiths et al., 2001; Kuan et al., 2007; Gregory et al., 2009). The effects may not be linear (Hirsch et al., 1993; Giller et al., 2009) as different groups dominate at different points as toxicity changes. Additions of Cu and Zn increase the metabolic quotient (Rost et al., 2001) which is an indicator of microbial stress (Chander & Joergensen, 2001; Chander et al., 2002). Metals added in association with sewage sludge may be more toxic than metals added as inorganic salts (Chaudri et al., 2008). As a result, contamination can decrease OM mineralization (Giller et al., 1998; Dai et al., 2004; Emmett et al., 2010), N fixation ( McGrath, 1998; Broos et al., 2004) and the catabolic diversity of soils, measured by the soil’s ability to degrade a range of compounds ( Wenderoth & Reber, 1999; Girvan et al., 2005). Urban soils often have less biodiversity (Fountain & Hopkin, 2004; Styers et al., 2010), presumably due to contamination.

Soil compaction. Compaction is mainly associated with trafficking in arable soils and livestock in grassland soils. Whereas clay soils are better able to recover, sandy soils are particularly vulnerable ( Gregory et al., 2009). Subsoil compaction is particularly insidious because it is rarely detected. Recent reviews have summarized comprehensively the causes, effects (some of which are discussed below) and management of soil compaction in the UK (Clarke et al., 2007; Batey, 2009).

Soil compaction degrades soil physical quality (Dexter, 1988; Whitmore & Whalley, 2009). Increases in bulk density (of up to 0.18 Mg/m, Ball et al., 2008) and soil strength (of up to 3 MPa, Gregory et al., 2007; Whalley et al., 2008), and decreases in friability (by up to 50%, Watts & Dexter, 1998), soil porosity (by 10–25%, Mooney & Nipattsuk, 2003; Gregory et al., 2009; Matthews et al., 2010) and water-holding capacity (Gregory et al., 2009) have been reported in UK soils. Compaction preferentially affects macropores (Brelan & Hansen, 1996; Richard et al., 2001) and can change the alignment of pores from vertical to parallel with respect to the soil surface (Servadio et al., 2005). This affects infiltration rates ( Heathwaite et al., 1990; Kibblewhite et al., 2008; Batey, 2009) and hence increases the risk of overland flow, flooding and erosion (Dexter, 1988). Decreases in the saturated hydraulic conductivity of soil by up to three orders of magnitude have been reported ( Mooney & Nipattsuk, 2003; Matthews et al., 2010). Compaction in grassland soils is less severe than in arable soil (Palmer & Smith, 2013), although it is much less understood (Bilotta et al., 2007).

Compaction may have little effect on micro-organisms as they inhabit small pores which are less affected (Brelan & Hansen, 1996; Jensen et al., 1996; Kohler et al., 2005; Shestak & Busse, 2005; Gregory et al., 2007). However, microbial functions may be affected. Brelan & Hansen (1996) reported reductions of up to 18% in N mineralization with compaction, ascribed to increased physical protection of SOM and microbial biomass in the smaller pores inaccessible to grazing nematodes created by compaction. Denitrification and NH3 volatilization may increase with compaction, causing N loss (Batey & Killham, 1986). Ball et al. (2008) found N2O emissions increased by up to 4 mg N/m²/h following compaction – twice that of the uncompacted control. In contrast, compaction may adversely affect soil macro-organisms. Reductions in Collembola (Heisler & Kaiser, 1995; Schrader & Lingnau, 1997), Arthropoda (arthropods) (Aritajat et al., 1977a,b) and Enchytraeidae (potworms) (Schrader et al., 1997; Rohrig et al., 1998), and both surface-dwelling (Pearce, 1984; Radford et al., 2001) and deep-burrowing Lumbricidae species (Rushton, 1986; Kretzschmar, 1991) have been reported.

Soil sealing. Soil sealing is associated with the main urban areas of the UK under the greatest pressure for housing and infrastructure. In 2011, 82% of the population of England lived in urban areas (ONS, 2012, 2013). Whilst there is little direct evidence of the impacts on soil properties in the UK, general impacts from a European perspective have been discussed (e.g. Huber et al., 2008; Scalenghe & Marsan, 2009; Virto et al., 2015). Sealing interrupts or removes completely the contact between the soil and other system components such as the biosphere, atmosphere and hydrosphere (Huber et al., 2008) with significant and irreversible impacts on the ability of soil to transmit water and gas (Virto et al., 2015). Declines in SOM (Wei et al., 2014) and microbial biomass and respiration (Piotrowska-Dlugosz & Charzyński, 2015) have also been reported.

Brownfield development. The impact from brownfield development on soil properties is wide-ranging as determined by the specific prior engineering project. In the extreme case where brownfield development results in a nongreen after-use, the soil is removed permanently and hence all soil functions will be lost. However, common brownfield
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Degradation threats to soil properties in the UK

From the review above, we see that the main degradation threats to soils in the UK affect a wide range of properties within the three main types: physical, biological and chemical. Soil erosion results in the removal of important topsoil and any contained nutrients, C and porosity. A decline in SOM principally affects soil biological and microbiological properties, SOM being the primary food source, but also impacts on soil physical properties because of the effect of SOM on soil structure development and stabilization. Soil contamination affects soil chemical properties, affecting nutrient availability and degrading microbial properties, whilst soil compaction degrades soil physical properties, especially soil porosity. Soil sealing removes the link between the soil and most of the ‘spheres’ and can significantly affect soil hydrological and microbial functions, but may not necessarily affect other soil properties. By contrast, soils on re-developed brownfield sites typically have degraded physical, chemical and biological properties. From our review, it is apparent that the potential or actual effect of some degradation threats on soils in the UK is well documented, including soil erosion, decline in SOM, soil contamination and soil compaction. By contrast, the impact of soil sealing and brownfield development on soils in the UK appears to be understood less. These areas might need to be research priorities in the future if the UK population (particularly in urban areas) continues to increase.

The spatial arrangement of degradation threats to soils in the UK is complex and beyond the scope of this review. However, we may offer some initial simplistic thoughts. Erosion is associated with overland flow on wet soils or wind in dry soils. Therefore, in the UK, upland soils in the north and west that experience considerable rainfall and arable soils of low porosity may be more prone to suffer from soil erosion by water, whereas light-textured arable soils may be more prone to wind erosion, such as those in the south and east of the UK. Intuitively, soils with a greater amount of SOM have more potential to lose SOM than soils of lesser SOM. As SOM is to a large extent dictated by land use, soils under grassland, semi-natural or natural vegetation may be candidates to experience a loss of SOM, particularly if the input from vegetation changes in some way. Grassland dominates soils in the west and north of the UK, whilst the main areas of semi-natural or natural vegetation are in the uplands. However, it is important to reiterate that there is currently no definitive evidence for significant changes to SOM levels in UK soils (Bellamy et al., 2005; Chapman et al., 2013; Reynolds et al., 2013; Gregory et al., 2014). Although compaction is often perceived to be a degradation associated exclusively with
trafficking of arable soils, grassland soils can become 
compacted from the impact of livestock, and hence, 
compaction can potentially affect many agricultural soils in 
the UK. The impacts are likely to be more severe in 
tramlines and where livestock congregate (e.g. feeding and 
watering troughs) in arable and grassland systems, 
respectively. Soil sealing and brownfield development are 
primarily urban-based degradations and hence will mainly 
threaten soils within and surrounding the main urban areas 
in the UK, particularly in the cities in England. Soil 
contamination could also affect the same soils close to urban 
and industrial areas, but other contaminants, particular 
airborne ones, could diffuse to affect soils distant from the 
source. Continued development and exploitation of soil and 
land-use spatial databases (e.g. Mackney et al., 1983; Proctor 
et al., 1998; Morton et al., 2011) should help to identify soils 
at risk from particular degradation threats. This kind of 
approach can then be used to review current land 
management practices in order to minimize degradation 
impacts.

**Implications for soil functions in the UK**

Having synthesized the literature on the impact of 
degradation threats on soil properties in the UK, we may 
then discuss the likely impacts on important soil functions – 
the tasks or services we ask of soil. These functions typically 
include food and fibre production, storage of water and C, 
water and air filtration, support for infrastructure, and 
support for biodiversity, habitats, cultural and archaeological 
heritage (HMG, 2011b). We discuss briefly some potential 
impacts of changes in soil properties arising from 
degradation threats on soil functions.

Loss of topsoil affects crop yields. A review of 24 studies 
in the UK found that yields decreased by 4% per 10 cm 
depth of soil loss, equivalent to a soil loss of around 100 
Mg/ha (Bakker et al., 2004). Erosion on lowland grasslands, 
which ranges from 0.5 to 1.2 Mg/ha, can be important 
locally for growth and silage quality (Bilotta et al., 2010). 
Erosion removes soil habitat space, thus impacting on 
biodiversity support and water storage functions. Amenity 
use is also affected by erosion (Rodway-Dyer & Walling, 
2010).

Degradation of soil physical properties, particularly the 
decrease in porosity from compaction and the destabilization 
of structure from loss of SOM, affects a range of soil 
functions. Water storage and flood regulation functions are 
primarily urban-based degradations and hence will mainly 
threaten soils within and surrounding the urban areas in 
the UK, particularly in the cities in England. Soil 
contamination could also affect the same soils close to urban 
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conditions, can severely impair plant growth and food quality (Pearson & Stewart, 1993; Blake et al., 1994; Blake & Goulding, 2002; Millennium Ecosystem Assessment, 2005; Degryse et al., 2007; Atkinson et al., 2012). Some metals and metalloids are phytotoxic and decrease yields at high concentrations (e.g. Zn, Cu, Cr and As), whereas ‘passage poisons’ affect animal and humans consuming food with little effect on yields (e.g. Cd, Mo, Se and possibly Co) (McGrath & Zhao, 2015). The UK has substantial areas that are contaminated with As (Appleton et al., 2009). Acidification can cause plant nutrient imbalances (Phoenix et al., 2004), increased risk of nutrient leaching (Pearson & Stewart, 1993; Jefferies & Maron, 1997; Degryse et al., 2007) and increased susceptibility of plants to stressors, including diseases and pests (Power et al., 1998; Carroll et al., 1999). Acid grassland can be toxic to grazing. Above-ground biomass may be significantly reduced near urban areas due, in part, to the proximity of contaminant sources (Ander et al., 2013). Ground-level O₃ may also potentially reduce grass (Gonzalez et al., 1999) and arable crop yields by up to 15% in the UK, in part by the effect on the ability of plants to respond to drought and to sequester C (ROTAP, 2009). Excess salt levels are toxic to most plant species, but the resulting halophytic ecosystems are often of great biodiversity value however (e.g. salt marsh, Watts et al., 2003). Indeed, soil pH is one of the main determinants of plant biodiversity (Bunemann et al., 2006; Cookson et al., 2006; Rousk et al., 2009).

At a fundamental level, where soil sealing removes the link between soils and the other ‘spheres’, functions such as supporting vegetation growth, C sequestration, filtering water and air and supporting biodiversity are significantly or irreversibly reduced (Huber et al., 2008; Scalenghe & Marsan, 2009). Also, the degradations in soil properties arising from brownfield development cause problems for recreating functioning soils. Restoration to a desired postoperation land use, such as species-rich grassland, may require intervention (Carrington & Diaz, 2011). However, modern restoration techniques can successfully recreate habitats of high nature conservation value (Tarrant et al., 2013), and indeed, soils with poor structure or nutrient status may help in this regard (HMG, 1996).

It is therefore apparent that if potential threats from soil degradation are realized, a range of important soil functions may be affected, through effects on measured soil properties. The studies reviewed above largely focus on functions of immediate current economic value (e.g. crop yields and water resource management) and were often empirical observations rather than mechanistic studies linked to soil properties. Extrapolation of knowledge relating soil properties and soil functions is challenging as a result (HMG, 2011b), even though similar processes may be at work. An associated issue is the potential or otherwise for soils to be truly multifunctional. There are obvious trade-offs: agricultural monoculture might be at odds with supporting biodiversity, for instance. In the future, a more mechanistic understanding of the response of soil functions to degradation processes, through measured changes in soil properties, would help in this regard.

It is important to note that degradation processes do not necessarily have irreversible effects on soil functions. Land management practices can be revised to reduce or even remove the degradation threat to improve soil properties and functions. Examples in agriculture include reducing compaction by cultivating soil in a brittle rather than plastic state as well as lowering tyre pressures on machinery (Batey, 2009) and decreasing erosion by tilling perpendicular to (rather than up-and-down) the slope (Quinton & Catt, 2004). Farmers may also offset SOM loss by incorporating residues or adding organic materials. On nonagricultural land, degradation threats such as contamination, sealing and brownfield site development can probably only be mitigated by policy.

The effects of climate change on soils in the UK

It is predicted that the UK will experience hotter, drier summers and warmer, wetter winters than currently, with an increased occurrence of low-frequency-high-magnitude events such as heat waves and intense storms (HMG, 2009a,b; Lowe et al., 2009). Climate change is likely to affect all the degradation threats, and hence soil properties. We briefly summarize some important implications with reference to observations in the literature and modelling. Whilst it may be possible to look at evidence from regions currently experiencing the kind of climate predicted for the UK, such soils have likely developed over millennia under stable conditions and hence may not necessarily be analogous to how UK soils will respond to rapid climate change.

Soils on the margins of stability are likely to be most vulnerable to drier summers, such as those in eastern England that already experience low rainfall, or peat soils that can suffer irreversible shrinkage upon drying. Prolonged drying can increase soil hydrophobicity (McHale et al., 2005) and strength (Whitmore & Whalley, 2009). Rapid drying can increase the mobility of cationic metals (Simpson et al., 2010), which can reduce the buffering capacity (Park et al., 2010). The likely effect on SOM is unknown: warming may increase microbial activity (Cox et al., 2002; Carney et al., 2007) or net primary productivity (Smith et al., 2007; Johnston et al., 2009), resulting in either a net loss or gain in soil C, respectively. Some models predict significant loss of soil C across much of the UK up to 2080 (Smith et al., 2005). Suseela & Dukes (2013) suggested a seasonality effect on SOM with decreased respiration during the growing season and increased respiration in the nongrowing season. Global warming is predicted to cause a sea-level rise of between 3 and 9 cm between 2010 and 2030 (Lowe et al.,...
which will increase the risk of flooding and salinization in low-lying coastal soils of the UK.

The wet winter experienced in the UK in 2000–2001 provided some insight into future rainfall patterns (ADAS, 2002). Soils suffered from prolonged waterlogging, rill erosion, losses of NO₃ (in drainage water), S (in SO₂) and P (in sediment), and precipitation of Fe and Mn (ADAS, 2002). Waterlogging causes anaerobism and has been found to reduce N, P and K nutrient uptake rates to <10% of control soils and to increase the concentration of potentially harmful compounds including C₂H₄ (up to 6 µL/L) and Mn (up to 0.5 mg/kg) (Drew & Sisworo, 1979). McTiernan et al. (2001) reported 70–300% greater losses of dissolved organic C from grassland that was seasonally waterlogged rather than well drained, although Scholefield et al. (1993) found increased NO₃ leaching from drained soil at the same site. Unger et al. (2009) found that flooding reduced total microbial biomass, Gram-positive bacteria and mycorrhizal fungi by up to 50%. Milan (2012) reported lasting effects of high-magnitude flooding in an upland catchment in the UK where soil adjacent to the river was lost together with its vegetation. The overall increase in flood frequency in the UK may be as great as 50% at the current 50-yr return period in some catchments (Kay et al., 2006). The full effects of significant flooding in southern England in the winter of 2013–2014 have yet to be assessed.

Changes in soil properties arising from climate change will affect important soil functions, particularly food and fibre yields. With increased drying, rainfall agriculture could become risky for some important crops in the UK such as potato (Daccache et al., 2012). Following the wet winter in the UK in 2000–2001, yields of winter wheat (12% decline), barley (7.5% decline), spring oilseed rape crop, sugar beet and potato were significantly reduced (ADAS, 2002). Rounsevell & Brignall (1994) concluded that machinery working days would be reduced should autumnal precipitation increase in England by 15%, which is entirely possible. Flooding makes it difficult to get livestock onto the land (Tyson et al., 1992), and animals grazing floodplains are at risk of being lost in floodwater (Blackwell & Maltby, 2006). While the greater atmospheric CO₂ may enhance yields of temperate plants (Long et al., 2004), it could come at a cost of reduced plant protein content (Cotrufo et al., 1998; DaMattia et al., 2010), if N becomes limited in the soil (Gill et al., 2002). Further impacts on soil functions may be apparent. Increased concentrations of CO₂ may reduce plant transpiration (Körner, 2000), which could reduce the ability of the soil to store additional rainfall. Substantially reduced flows in wetlands may reduce their water regulatory capacity (Sutherland et al., 2008). Waterlogging causes reducing conditions in soils, which could impact on buffering and contaminant filtering (Du Laing et al., 2009; Lair et al., 2009). The impact of climate change on soil biodiversity remains unclear, although it will reflect changes in general soil conditions and vegetative growth. Resilience in biodiverse habitats at the larger scale may mask changes at fine scales (Fridley et al., 2011). A link between climate, shrink–swell cycles in clay soils and subsidence claims from properties damaged in the UK has been reported (Harrison et al., 2012), and it is predicted that the magnitude and frequency of landslides in the UK will increase in the future (Pritchard et al., 2014). There is conflicting evidence as to whether waterlogging and anaerobism will help to preserve (Caple, 1996; Raiswell, 2001) or degrade (Douterelo et al., 2010) archaeological materials in soils.

It would appear that soils will need to be carefully managed in the future to minimize any exacerbating effects of climate change on the range of degradations that they currently experience.

Conclusions

With the UK as an example, we have summarized some key data on the impact of degradation threats on soil properties and have sought to link changes in soil properties to important soil functions. The key next stage will be to continue to add to the database with field-based monitoring of broad changes to soil properties following degradation supported by controlled laboratory experimentation to improve understanding of the mechanisms involved. Outputs from such an approach, together with refinement of soil and land-use spatial databases, should aid the development of models capable of extrapolating up to the national scale in order to link soil degradations to soil functions. We may also be able to predict any effect that climate change is likely to have. Armed with this, we may revise our policies so that soils are truly managed sustainably.

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