Resilient remediation: Addressing extreme weather and climate change, creating community value

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Abstract  
Recent devastating hurricanes demonstrated that extreme weather and climate change can jeopardize contaminated land remediation and harm public health and the environment. Since early 2016, the Sustainable Remediation Forum (SURF) has led research and organized knowledge exchanges to examine (1) the impacts of climate change and extreme weather events on hazardous waste sites, and (2) how we can mitigate these impacts and create value for communities. The SURF team found that climate change and extreme weather events can undermine the effectiveness of the approved site remediation, and can also affect contaminant toxicity, exposure, organism sensitivity, fate and transport, long-term operations, management, and stewardship of remediation sites. Further, failure to consider social vulnerability to climate change could compromise remediation and adaptation strategies. SURF’s recommendations for resilient remediation build on resources and drivers from state, national, and international sources, and marry the practices of sustainable remediation and climate change adaptation. They outline both general principles and site-specific protocols and provide global examples of mitigation and adaptation strategies. Opportunities for synergy include vulnerability assessments that benefit and build on established hazardous waste management law, policy, and practices. SURF’s recommendations can guide owners and project managers in developing a site resiliency strategy. Resilient remediation can help expedite cleanup and redevelopment, decrease public health risks, and create jobs, parks, wetlands, and resilient energy sources. Resilient remediation and redevelopment can also positively contribute to achieving international goals for sustainable land management, climate action, clean energy, and sustainable cities.

Keywords  
climate change, extreme weather events, remediation resiliency, sustainable remediation, sustainable remediation forum

1 | INTRODUCTION

After Hurricane Harvey made landfall in 2017, 13 Superfund sites in Houston, Texas, were flooded. At one site, the U.S. Environmental Protection Agency (U.S. EPA) measured dioxin at levels over 2,300 times the level requiring cleanup actions. Five weeks after Hurricane Maria hit in 2017, one in four Puerto Ricans lacked access to clean water. During Hurricane Florence, U.S. EPA and scientists from industry, universities, and civil organizations warned of the potential release of toxic chemicals from North Carolina and South Carolina Superfund sites. Post landfall, Florence led to extensive flooding that “… swept away part of a retaining wall holding back a pond of coal ash – which contains mercury, arsenic and other toxic substances – and have also overrun several lagoons of pig waste in North Carolina” (Pierre-Louis, Popovich, & Tabuchi, 2018, p. 1).

In the United States, nearly two million people—the majority in low income communities—live within 1 mile of one of 327 Superfund sites in areas prone to flooding or vulnerable to sea-level rise caused by climate change (Dearen, Biesecker, & Kastanis, 2017). These 327 sites are part of a much larger universe of U.S. sites that need to be assessed. There are more than 650,000 contaminated commercial and industrial sites and more than 81,000 acres of
brownfields at 21,000 sites in 232 cities across the United States (Targ, 2017).

Globally, the number of contaminated sites is overwhelming and growing as a result of increasing urbanization especially in emerging economies. Estimates for Europe alone (excluding many diffuse land contamination problems) range from 2.5 to 4.5 million sites. In China, about 20% of farmland is contaminated by trace metals, pesticides, and hydrocarbons such as petrochemicals (Bardos, Bakker, Slenders, & Nathanail, 2011). Over one million contaminated sites may require cleanup (Hou & Li, 2017), nearly 60% of groundwater is not safe for drinking (Hou, Li, & Nathanail, 2018), and public health threats can exist even on contaminated land that has been remediated.

2 | RESEARCH FINDINGS

Decades of research, including the recent 2017 U.S. Climate Science Special Report (Wuebbles et al., 2017), document the global reality of more powerful and frequent storms, heavy rainfall, heat waves, wildfires, and more frequent and longer droughts. Rising sea levels, declining snowpack, long-term stress on water availability, dynamic groundwater levels, acidification, and rising temperatures represent further threats to ecosystems and communities.

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At hazardous sites, climate change and extreme weather events can undermine the effectiveness of the original site remediation design and can also impact contaminant toxicity, exposure, organism sensitivity, fate and transport, and long-term operations, management, and stewardship of remediation sites.

Higher temperature and lower pH can increase the availability of contaminants in the environment. For example, the speciation and availability of metals changes with environmental pH (Millero, Woosley, DiTrioio, & Waters, 2009), and the fate and transport of persistent organic pollutants changes with temperature and precipitation (Nadal, Marques, Mari, & Domingo, 2015).

Increasing temperatures can also change the water cycle influencing the local water budget. Warmer temperatures can result in altered precipitation, increased evaporation rates of surface water, increased rates of water uptake by vegetation, and reduced rates of water recharge to soils and groundwater reservoirs (Famiglietti, 2014).

Increased temperatures and changes to the water cycle may also result in more frequent and severe weather events, such as the occurrence of the 100-year storm event, as well as, contribute to more frequent nuisance flooding due to the prevalence of supersaturated soils. Both events are exacerbated by sea level rise resulting in shoreline encroachment and increased nuisance flooding during high tide.

Additional vulnerabilities of water resources include, but are not limited to, changes to water supplies, subsidence, increased amounts of water pollution, erosion, and related risks to water and wastewater infrastructure and operations, degradation of watersheds, alteration of aquatic ecosystems and loss of habitat, creating multiple impacts in coastal areas (LARWQCB, 2015). These hydrological changes are happening at the same time as groundwater extraction is increasing as heat also increases demand for various water needs, including drinking, irrigation, and industrial uses (Famiglietti, 2014).

A recent study showed a potential impact of such climatic shifts on residual contaminants in soil and groundwater (Libera et al., 2018). The study found that the hydrological shifts influence contaminant concentrations in a complex manner, since increased infiltration, for example, could cause conflicting effects of both diluting and mobilizing contaminants. The study showed that, in general, higher infiltration events could mobilize vadose-zone residual contaminants, raising contaminant concentrations in groundwater for a prolonged period.

Similarly, the sensitivity of organisms and ecosystems can be affected by environmental change. Higher temperatures increase the metabolic rate of ectotherms (organisms that derive their heat and, therefore, maintain their metabolic activity from the environment around them), which can increase the rate at which they absorb or process contaminants (Noyes et al., 2009). Behavioral changes in response to environmental change may also alter exposure and sensitivity as organisms react to new stresses in ways that ameliorate or exacerbate other stresses.

The use of the chemicals that become environmental contaminants is also likely to change. For example, warming temperatures leads to expansion of agricultural pests, resulting in increased use of pesticides. Furthermore, more rain may require repeated application of pesticides and fertilizers. Both scenarios can result in agricultural land contaminated by intense application of chemicals as well as contributions to polluted runoff that impact nearby and downgradient waterbodies.

Climate change also poses challenges for selecting remediation techniques, including the feasibility of passive remediation technologies (O’Connell & Hou, 2015). Passive remediation carries an increased burden of proof, since contaminants stay longer in the subsurface—compared to conventional soil removal options—while degradation/treatment processes occur.

Thus, the efficacy of remediation efforts may be undermined if attention is not paid to climate change impacts throughout the remediation process. This can be thought of from two different perspectives: (1) how climatic change will affect remediation, and (2) how remediation techniques will be affected by climate change. Examples of each are presented in Exhibits 1 and 2.
### EXHIBIT 1  Implications of climate change for remediation

<table>
<thead>
<tr>
<th>Climate impact</th>
<th>Secondary effect</th>
<th>Relevant remediation effect</th>
</tr>
</thead>
<tbody>
<tr>
<td>Altered precipitation pattern</td>
<td>Wetter: Flooding, storms, more runoff</td>
<td>Mobilization of contaminants (e.g., from vadose zone to groundwater) → Higher contaminant concentration/export, overpowering significant degradation rate in groundwater zone could remove natural protective barriers or cause infill subsidence in low-lying areas</td>
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<tr>
<td></td>
<td>Drier: Drought</td>
<td>Damage to capping systems</td>
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<tr>
<td></td>
<td></td>
<td>Dilution → Lower contaminant concentration/export</td>
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<tr>
<td></td>
<td></td>
<td>Damage to capping systems</td>
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<td></td>
<td></td>
<td>Oxidation of soils</td>
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<td></td>
<td></td>
<td>Increased volatility</td>
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<td></td>
<td></td>
<td>Less dilution → Higher contaminant concentration/export</td>
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<tr>
<td></td>
<td></td>
<td>Reduced mobilization → Higher contaminant persistence (higher contaminant concentration/export)</td>
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<tr>
<td></td>
<td></td>
<td>Insufficient water for remediation; Overuse of groundwater</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Possible enhanced natural attenuation, expedited contaminant removal</td>
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<tr>
<td></td>
<td>Altered salinity</td>
<td>Altered degradation rates (physical, microbial)</td>
</tr>
<tr>
<td>Sea level rise</td>
<td>Erosion</td>
<td>Damage to site integrity</td>
</tr>
<tr>
<td></td>
<td>Site inundation</td>
<td>Increased mobilization of contaminants, possible dilution, or compromised site with mixing or loss of contaminated materials, increased bioavailability of contaminants</td>
</tr>
<tr>
<td></td>
<td>Mobilization of contaminants</td>
<td>Clean sediments transported on top of contaminated sediments</td>
</tr>
<tr>
<td></td>
<td>Elevations increase</td>
<td>Changing footprint of flood plains, river boundaries, and coastal shoreline encroachment → Impact on regulations (e.g., dredging, cleanup levels, negotiation of water levels, monitoring)</td>
</tr>
<tr>
<td>Extreme weather</td>
<td>Scour (wind/wave action; surface water flow velocity)</td>
<td>Damage to site integrity, capping systems</td>
</tr>
<tr>
<td></td>
<td>Flooding</td>
<td>Possible dilution (lower contaminant concentration/export), or compromised site with mixing or loss of contaminated materials, damage to capping systems</td>
</tr>
<tr>
<td></td>
<td>Extreme heat</td>
<td>Increased volatility → Mobilization of contaminants from site through soil and air</td>
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<tr>
<td></td>
<td>Freezing conditions</td>
<td>Changes in use of site by wildlife</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Melting permafrost → Mobilization of contaminants from site through water, soil, and air</td>
</tr>
<tr>
<td>Extreme weather: Fire</td>
<td>Increased use of fire retardants</td>
<td>Damage to site integrity, capping systems</td>
</tr>
<tr>
<td></td>
<td>Damage to site infrastructure</td>
<td>Spread of contaminants</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Loss of function of remediation systems</td>
</tr>
<tr>
<td>Decreasing pH</td>
<td></td>
<td>Increased availability, mobilization, toxicity</td>
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<tr>
<td></td>
<td></td>
<td>Increased sensitivity of species due to pH stress</td>
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<tr>
<td></td>
<td></td>
<td>Altered transformation rates</td>
</tr>
<tr>
<td>Increasing temperature</td>
<td>Altered transformation or degradation</td>
<td>Increased or decreased toxicity</td>
</tr>
<tr>
<td></td>
<td>Decreased dissolved oxygen/anoxic conditions</td>
<td>Altered transformation, decreased species resilience</td>
</tr>
<tr>
<td></td>
<td>Increased species heat stress and associated conditions</td>
<td>Increased sensitivity to contaminants</td>
</tr>
<tr>
<td>Human impact and responses</td>
<td>Vulnerable communities commonly comprised of low socioeconomic and minority populations</td>
<td>Cardio-pulmonary illness</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Food, water, and vector-borne diseases</td>
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<tr>
<td></td>
<td></td>
<td>Loss of homes, drinking water, and livelihoods</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Mental health consequences and stress</td>
</tr>
<tr>
<td></td>
<td>Increased use of some chemicals</td>
<td>Additional toxicity, additional remediation sites.</td>
</tr>
<tr>
<td></td>
<td>Conflicting solutions, changing land use demands, shifting populations</td>
<td></td>
</tr>
</tbody>
</table>
EXHIBIT 2    Impact of climate change on remediation techniques*

<table>
<thead>
<tr>
<th>Remediation approach</th>
<th>Technique</th>
<th>Climate change impact</th>
</tr>
</thead>
<tbody>
<tr>
<td>Soil treatment</td>
<td>Bioremediation</td>
<td>Degradation activity may change, unexpected intermediaries</td>
</tr>
<tr>
<td></td>
<td>Land farming/landspreading</td>
<td>Inundation of site by sea level rise or flooding</td>
</tr>
<tr>
<td>Groundwater treatment</td>
<td>Pump and treat</td>
<td>Altered rate of recharge and extraction</td>
</tr>
<tr>
<td>Removal of contaminated materials</td>
<td>Soil washing</td>
<td>Extreme weather, flooding, or sea level rise will complicate containment</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Groundwater level decline may support expedited removal</td>
</tr>
<tr>
<td>Engineered in situ solutions</td>
<td>Soil extraction</td>
<td>Insufficient water would limit feasibility</td>
</tr>
<tr>
<td></td>
<td>Natural attenuation</td>
<td>Warmer temperatures may help</td>
</tr>
<tr>
<td></td>
<td>Incineration</td>
<td>Models do not include climate change which may alter resident time of contaminants in soil → Attenuation rates may vary</td>
</tr>
<tr>
<td></td>
<td>Capping systems</td>
<td>Emissions allowances may change due to temperature or greenhouse gases</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Climate change may degrade the cap (e.g., because of extreme precipitation events) → Much higher contaminant concentration/export and increased mobilization of contaminants in vadose zone</td>
</tr>
</tbody>
</table>

*Note. See also U.S. EPA fact sheets developed for contaminated sediments, groundwater remediation systems, and landfills and other containment remediation.

3 | SOCIOECONOMIC AND LEGAL IMPLICATIONS

The National Climate Assessment (NCA) provides an in-depth assessment of climate change impacts on the lives of Americans; the Fourth NCA noted that "extreme weather events have cost the United States $1.1 trillion since 1980" (Hibbard, Dokken, Stewart, & Maycock, 2017); and The U.S. Government Accounting office warned that climate change "could increase flooding costs in coastal communities by $23 billion per year by midcentury" (Plumer, 2017).

Communities adjacent to contaminated sites are often comprised of socioeconomically depressed and environmental justice (sensitive) populations that usually have little influence over the decision-making process, even when they are most impacted. North Carolina residents evacuated during Hurricane Florence shared the fate of New Orleans residents post Katrina “…the poor are always vulnerable- to the perceived values of their residences in good times and the ravages of Mother Nature when disaster hits” (Fausset, 2018, p. 1).

Parties liable under the U.S. Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) can face additional liability if global warming-related weather events exacerbate problems on contaminated properties. There is no minimum quantity of a hazardous substance needed to establish liability, and a generator or transporter is liable whether or not the hazardous substances they generated or transported are the primary contaminants of concern at the site. All of the parties (current and past owners and operators, generators, and transporters) are also liable if contaminants migrate from the original disposal area.

CERCLA contains an “Act of God” defense, defining an “Act of God” as “an unanticipated grave natural disaster or other natural phenomena of an exceptional, inevitable, and irresistible character, the effects of which could not have been prevented or avoided by the exercise of due care or foresight” (42 U.S.C. § 9601(1)(1980)). Second, the defendant will have to prove the event was “unanticipated” (42 U.S.C. § 9601(1)(1980)). Third, the defendant will have to prove that the effects of the event “could not have been prevented or avoided by the exercise of due care or foresight” Id. The failure to date of the Act of God defense is illustrated by the results in the cases in which it has been unsuccessfully attempted (see, e.g., U.S. v. Stringfellow, 661 F.Supp. 1053 (C.D. CA 1987); U.S. v. W.R. Grace & Co.-Conn., 280 F.Supp.2d 1135 (D. MT 2002); U.S. v. Alcan Aluminum Corp., 892 F. Supp. 648 (M.D. Pa. 1995), aff’d, 96 F.3d 1434 (3d Cir. 1996); U.S. v. Barrier Industries, Inc., 991 F. Supp. 678 (S.D. N.Y. 1998); U.S. v. M/V Santa Clara I, 887 F. Supp. 825, 843 (D.S.C. 1995)).

As part of a U.S. government-wide effort, the U.S. EPA began analyzing how climate change could impact the nation’s most hazardous sites and developing best practices for the most successful remediation techniques (https://www.epa.gov/superfund/superfund-climate-change-adaptation). U.S. EPA also reported on additional community benefits of climate change adaptation at “brownfields” and recommended land use, zoning, and building code changes and/or development incentives that could increase resiliency (https://www.epa.gov/brownfields/climate-adaptation-and-brownfields).

3.1 | Overarching resilient remediation principles

Sustainable Remediation Forum (SURF) recommendations to advance climate change resilience within contaminated lands rehabilitation build on these U.S. EPA initiatives, along with well-established climate change adaptation tenets, marrying them with sustainable remediation principles and practices. SURF’s Sustainable Remediation Framework calls for “a systematic, process-based, iterative, holistic approach beginning with the site end use in mind” (Holland et al., 2011). This holistic approach can incorporate planning for uncertainty, reducing the rate and extent of local, regional, and global climate change impacts, and address social impacts, equity concerns, and opportunities. Setting criteria and indicators for measuring progress provide for more transparency and can gain stakeholder support.
While the toxicological literature includes a fair amount of understanding regarding how the parameters related to climate change (temperature, pH, salinity, dilution) affect contaminants, there is little application of these parameters in combination (as will often be the case with climate change). As a result, it will be necessary to develop approaches to remediation that can be adapted as new information is gathered in the treatment process.

To be most effective, adaptation should be an iterative and flexible process that involves periodically reevaluating the remediation system’s vulnerability, monitoring the measures already taken, and incorporating newly identified options or information into the adaptation strategy. This involves consideration of short- and long-term availability of resources, such as energy and clean water, and ecosystem services as well as land uses of site or the surrounding area that may be critical aspects of the remediation system (U.S. EPA, 2015). As part of this iterative and flexible process site managers can use scenario planning that details future potential conditions in a manner that supports decision-making under conditions of uncertainty but does not predict future change that has an associated likelihood of occurrence (Glick, Stein, & Edelson, 2011).

Considering the role of remediation in greenhouse gases (GHGs) emissions is important. Energy-intensive remedies are often a significant source of GHGs. At one remediation project in New Jersey, it was estimated that the difference between two proposed remedies could be as high as 2% of the annual GHG emissions for the entire state (Ellis & Hadley, 2009). Further, a meta-analysis indicated that the cleanup of 1 kg of contaminants in groundwater may result in up to 130 tons of CO₂ emissions, with a geometric mean of 1.3 tons of CO₂ emissions (Hou & Al-Tabbaa, 2014).

As part of the sustainable remediation assessment, these GHGs determinations can support decisions that reduce the manifestations of climate change on the site. Best management practices can be found in the ASTM Greener Cleanups and Consideration Sustainability in Remediation Projects guidance documents.

Social vulnerability is an ability to cope with and adapt to any external stress placed on livelihoods and well-being (Adger & Kelly, 1999). Adaptation strategies need to identify stakeholder concerns and address risk perception barriers. These strategies can include localized investigation to find answers to the questions about whom and what are vulnerable, to what are they vulnerable, how vulnerable are they, what the causes of their vulnerability are, and what responses can lessen their vulnerability (U.S. National Oceanic and Atmospheric Administration [NOAA], 2016; NOAA Community Social Vulnerability Indicators [Coburn, Spence, & Pomonis, 1994]). For example, the local community, municipal planners, and office of emergency management representatives can inform site managers on areas in and within the vicinity of the site that experience frequent nuisance flooding and are vulnerable to severe weather events.

Strategies for resilient rehabilitation of contaminated sites should:

- Involve the community throughout the cleanup and redevelopment process.

### 3.2 Site-specific protocols

Aligned with these overarching principles are recommended site-specific protocols that begin with the U.S. EPA and ASTM guidance and recent Washington State guidance, Adaptation Strategies for Resilient Remedies (Washington State Department of Ecology [DOE], 2017). The Washington DOE guidance is intended to: (1) help understand site-specific vulnerabilities of cleanup sites to climate change impacts, and (2) provide recommendations to increase resilience of remedies at each phase of cleanup. The guidance focuses on four climate change impacts: sea-level rise, flooding, landslides, and wildfires. The Washington DOE guidance also includes examples of vulnerability analyses, a list of references, links to different technologies, adaptation plans, decision tools, case studies, and sustainable remediation resources.

Exhibit 3 depicts U.S. EPA’s climate vulnerability and adaptation model, which evaluates the sensitivity, exposure, and adaptive capacity of the site, contaminant, or remediation technique to climate change.

An evaluation of a system’s vulnerability to climate change involves identifying climate change hazards of concern (such as treatment or containment systems) in light of potential climate/weather and considering factors that may exacerbate the system’s exposure and sensitivity, such as a long operating period.

For riverine and coastal sites, a vulnerability assessment may also encompass hydraulic and hydrological modeling or hydrodynamic modeling, respectively, to evaluate the role of precipitation projections, storm surge, surface water flow velocity, sea level rise, wave...
action, and/or wind action under existing and future storm events (e.g., 100-year storm event). The results of the modeling aid in remedial design, such as armor stone specifications for cap enhancement, and periodic climate change vulnerability monitoring, such as continuous monitoring of water levels, wave action, and flow velocity.

To support these vulnerability assessments, practitioners should use best available guidance (e.g., see Exhibit 4) and confer with local/regional experts and affected communities. Dynamic geospatial data are available from several sources, including federal, state, regional, or local sources such as watershed and forestry management authorities, nonprofit groups, and academia.

The site vulnerability assessment process should involve local government and residents: Stakeholder engagement strategies can include focus group interviews, local workshops, and/or public comment periods. This process can also increase local understanding of the risk of climate change and provide new perspectives on remediation options (Harclerode et al., 2015).

The vulnerability assessment should identify the need for adaptation strategies and long-term vulnerability monitoring protocols as part of operation and maintenance (O&M).

Adaptation strategies can also leverage existing regulatory tools such as the NCP long-term effectiveness and permanence (40 CFR 300.430(e)(9)(iii)(C)) and the Superfund Five-Year Reviews (Thun, 2017).

Five-Year Reviews should include the following elements (Thun, 2017):

- Evaluate remedy implementation/performance to determine protectiveness.
- Determine if the remedy functioning as intended. QUESTION C: Has any other information come to light that could call into question the protectiveness of the remedy?
- Address site changes or vulnerabilities that may be related to climate change impacts not apparent during remedy selection, remediation, or O&M (e.g., sea level rise, changes in precipitation, increasing risk of floods, changes in temperature, increasing intensity of hurricanes and increasing wildfires, melting permafrost in northern regions, etc.).
- Determine if the assumptions, data, and cleanup levels are still valid and, if there are issues, update O&M or remedy decision.

## 4 ADAPTATION STRATEGIES CASE STUDIES

Adaptation strategies can be categorized as resistance, resilience, and response.

Resistance strategies maintain current conditions. They can include physical security, such as hardening covers, caps, and barriers to prevent flooding or erosion. Resistance strategies eventually will succumb to change or need to be increased at continuing cost.

Resistance strategies allow sites to experience the change but still manage contaminant mitigation successfully. For example, to improve protectiveness and long-term effectiveness against more frequent severe storms, damaged portions of an intertidal cap at the Port Gamble Bay and Mill Site in Kitsap County, Washington, were repaired and replaced with armor of rocks and other natural materials almost twice the original size (Washington DOE, 2017).

Resilience strategies also include backup power and remote and communication including automated data acquisition. An example developed by the Lawrence Berkeley Laboratory for the Department of Energy capitalizes on 21st century technology through a new streamlined real-time data processing and analysis and early warning system for the Savannah River Superfund Site F-Area with a 50% cost saving (Exhibit 5, Wainwright, 2016).

Resilience strategies can also include the use of recycled water, including treated groundwater, to respond to drought conditions or salt water intrusion.

Another example of resilience comes from Huangshi in south central China, where intensive mining and smelting have caused significant air and water pollution and the contamination of nearby agricultural lands. Strip mining resulted in over 100 man-made bluffs, which are susceptible to landslides. One of the Rockefeller Foundation 100 resilient cities, Huangshi helped stabilize the land at these abandoned sites to prevent flooding and protect resources and human health. These efforts included controlling water pollution through sewage collection, water treatment, and increasing vegetation with ecological restoration projects.

Response strategies range from pre- and post-site inspection to removal of some or all of the contamination. For example, the New Jersey Department of Environmental Protection (NJDEP) developed response strategy guidance targeted to site owners and persons responsible for conducting and overseeing cleanup (i.e., “Licensed Site Remediation Professionals”). After storms, all sites should be reevaluated to determine if any immediate environmental concerns needing action arose and whether site conditions changed requiring reassessment (NJDEP, 2016).
Responsible parties and regulators employed another effective response strategy at the Purity Oil Sales Superfund Site in Fresno, California. Over a period of 5 years, drought and agricultural pumping caused the groundwater table to drop more than 16 feet. The parties agreed to remove contamination from the newly exposed vadose zone through soil vapor extraction (SVE). SVE expedited the cleanup and prevented further migration of contaminants to groundwater, removing contamination orders of magnitude greater than more traditional pump-and-treat systems (Dailey, 2016).

Another example of the impact of extreme weather and heavy precipitation, and the vital importance of adequate response strategies come from Japan. Radionuclides from the Tokyo Electric Power Company Fukushima Daiichi nuclear power plant accident were released into the atmosphere and then deposited on land and sea surfaces. The government-commissioned decontamination work at the plant from 2011 to 2017 generated approximately 20 million cubic meters of removed contaminated soil. Most of the soil was stored in approximately 1,000 temporary storage facilities. Transportation of the soil to interim storage facilities started in 2015, and about 80% of the contaminated soil is still in the temporary storage sites.

Heavy rainfall in September 2015 caused torrential rains and flooding in the Kanto and Tohoku region in Japan, and the outflow of 448 of these temporary containers on agricultural land along two rivers. Emergency responders collected almost all the containers (five were left in the places inaccessible to the public and repaired). As follow-up, the Japan Ministry of the Environment developed guidelines, “Implementation of Appropriate Initial Response” for dealing with challenges associated with the storage of contaminated soil (Japan Ministry of the Environment, 2015). For example, when disasters are predicted, storage areas need to be checked in advance, and parties need to implement an emergency response plan to minimize the damage of contaminant releases.

Post Florence, “In collaboration with state partners and once conditions allow” the U.S. EPA committed to deploy Superfund Reconnaissance Teams to conduct visual inspections of affected site, document site conditions, potential migration of contaminants, and restoration of utilities (if applicable), and complete the field survey checklist and photographs (U.S. EPA, 2018a, p. 1).

In addition to the Washington and New Jersey initiatives highlighted above, Massachusetts and California have also established noteworthy programs as described later. Massachusetts enacted legislation (Green Communities Act and Global Warming Solutions Act) that provides rigorous clean energy goals designed to grow its clean energy economy, increase its energy independence, and reduce the pollution that contributes to climate change. The Massachusetts governor also issued an executive order establishing an Integrated Climate Change Strategy. The Massachusetts Department of Environmental Protection (2014) promotes the use of “greener cleanup” principles and practices for the assessment and remediation of oil and hazardous material disposal sites through regulation and guidance, and is evaluating regulated sites and their vulnerability to climate change impacts through a statewide Geographic Information System (Potter, 2017).

California’s Climate Adaptation Strategy can be leveraged to address climate resilience of contaminated lands, such as: (1) decarbonized (40% GHG reduction from 1990 levels by 2030), decentralized energy (50% Renewables Portfolio Standard by 2030), and (2) protection of the most vulnerable communities through the Sustainable Communities and Climate Protection Act linking GHG reduction efforts to transportation and land planning requirements (California Climate Adaptation Strategy; California Natural Resources Agency, 2009).

The State Water Resources Control Board Resolution #2017-0012: Comprehensive Response to Climate Change provides support for drinking water systems and disadvantaged communities, and improves ecosystem resilience in response to the effects of climate change (California State Water Resources Control Board, 2017). Further, the Los Angeles Regional Water Quality Control Board Framework for Climate Change Adaptation and Mitigation (2015) looks at the impact of effects of climate change on contaminated sites and underground storage tanks and how these effects can be taken into account in the Regional Water Board’s actions.

Finally, the California Department of Toxic Substances Control is developing climate change guidance specific to hazardous waste treatment, storage and disposal facilities, and the cleanup of contaminated sites.
5 CLIMATE RESILIENT REDEVELOPMENT: DRIVERS AND CASE STUDIES

Over the last decade U.S. and European Union (EU) initiatives have sought to advance remediation by assessing the benefits of rehabilitated land in strengthening community, economic, and ecosystem resilience.

A U.S. EPA 5-year study of brownfields found that residential property values increased from 5.1 to 12.8% after a nearby brownfield was assessed or cleaned up (Bartsch, 2015). The study also determined that brownfields cleanup can increase overall property values within a 1 mile radius by $0.5 to $1.5 million. In 2016, U.S. EPA also published guidance regarding brownfield Revitalization in Climate-Vulnerable Areas including ordinance regulation and development incentives (U.S. EPA, 2016).

Working for the City of San Francisco, Hou, Song, Hou, O’Connor, & Harclerode (2018) developed a method based on life cycle assessment of GHG emissions to compare brownfields to greenfield land development. The team examined three categories: (1) primary impact (associated with physical state of brownfield sites and greenfield sites), (2) secondary impact (associated with remediation activities at brownfield sites), and (3) tertiary impact (associated with post-remediation usage of the brownfield sites and avoided usage of greenfield land). Overall, the results show that the City’s brownfield land redevelopment could lead to a net GHG reduction of 51.9 million metric tons (Mt) CO₂ eq. over a 70-year period, or 0.74 Mt CO₂ per yr, the equivalent of 14% of San Francisco’s GHG emissions in 2010.

Overall, the results show that the City’s brownfield land redevelopment could lead to a net GHG reduction of 51.9 million metric tons (Mt) CO₂ eq. over a 70-year period, or 0.74 Mt CO₂ per yr, the equivalent of 14% of San Francisco’s GHG emissions in 2010.

The RE-Powering America’s Land Initiative, where U.S. EPA supports renewable energy development on potentially contaminated land, landfills, and mine sites, tracks the economic and environmental benefits associated with completed sites. Common benefits reported from developers/public agencies include revenues from land leases and taxes, electricity cost savings, job creation, and reduced GHG emissions (U.S. EPA, 2018b).

A recently completed renewable energy project in the San Francisco Bay Area, the Marin Clean Energy (MCE) Solar One partnership, exemplifies the RE-Power America benefits.

MCE Solar One repurposed 60 acres of a remediated brownfields site leased by Chevron to MCE Solar One for $1 per year. At 10.5 megawatts, MCE Solar One will eliminate 3,234 Mt of carbon dioxide in one year, equivalent to taking more than 680 cars off of the road annually. MCE Solar One provided community benefit by partnering with RichmondBUILD, a public–private partnership that focuses on training for skilled construction, hazardous waste removal, and renewable energy jobs. All RichmondBUILD participants come from low-income households. In addition, almost $2 million dollars was spent on project materials purchased or rented locally. The project also includes an innovative procurement approach called “community choice energy,” in which a public agency offers citizens and businesses an alternative to the utility for purchasing their electricity. As a result of the MCE Solar One project, homes and businesses now benefit from a more renewable electricity option that costs 2 to 5% less than the traditional Bay Area utility rates (https://www.mcecleanenergy.org/news/press-releases/mce-solar-one-thinking-globally-building-locally/).

6 EU AND UK DRIVERS AND CASE STUDIES

There is presently a trend across Europe for densification as a planning approach for sustainable development to foster efficient use of resources, efficient transport systems, and a vibrant urban life (e.g., Haaland & van den Bosch, 2015). Development often takes place on areas that are often viewed as underutilized land (such as green space, marginal land) or through redevelopment on previous industrial estates (dereelict, brownfield sites). However, this approach has also been challenged for its threat to urban green spaces (Haaland & van den Bosch, 2015) since together with urban Brownfields they potentially have an important role for offering climate change adaptation solutions.

This is strongly related to a much wider European debate about “Nature Based Solutions” (NBS), their importance in urban areas and how they might be managed and, if necessary, regenerated. The concept of NBS was introduced toward the end of the 2000s by the World Bank (MacKinnon, Sovrvelia, & Hickey, 2008) and International Union for Conservation of Nature (IUCN, 2009) to highlight the importance of biodiversity conservation for climate change mitigation and adaptation. NBS were proposed by IUCN for inclusion in the climate change negotiations in Paris “as a way to mitigate and adapt to climate change, secure water, food and energy supplies, reduce poverty and drive economic growth” (IUCN, 2014). The IUCN proposed principles for NBS included cost efficiency, harnessing both public and private funding, ease of communication, and replicability of solutions (van Ham, 2014). Thus, NBS puts an explicit emphasis on linking biodiversity conservation with goals for sustainable and climate resilient development (Eggermont et al., 2015), and represent innovative, implementable "solutions."
The Holistic Management of Brownfield Regeneration (HOMBRE) was a major EU project completed in 2014 (www.zerobrownfields.eu), examining the enhanced transition of brownfields through to becoming once more a functional part of the land cycle. One of its areas of interest was in “soft,” that is, non-built reuse of brownfields, the services this might provide, and how those might be appreciated and valued. One of the outputs of this work is a simple Excel design aid to help developers and others involved in brownfields map the range of opportunities, the resulting value, and the initial design considerations by identifying specific opportunities for synergies between different “services” such as risk management, water improvement, and renewable energy.

6.1 | Case study: Brownfields redevelopment as wetlands park and community management

Port Sunlight Riverside Park: Port Sunlight River Park is a 28-hectare park near Birkenhead in Wirral, Merseyside, UK, which opened in 2014. It is located on a former landfill capped and covered by the waste management company (Biffa Waste Management) and leachate and gas management systems were put in place. The site was passed over to the Land Trust on a 99-year lease and, after planning and design, was created as a riverside park in 2013. The waste management company remains responsible for ongoing management and monitoring of the capping, landfill gas, and leachate treatment.

The Land Trust secured a £3.4 million investment for a transformation project encompassing park creation, site of special protection and ongoing management, and established a partnership with the charity, Autism Together, which manages the park.

A retrospective qualitative sustainability assessment was performed by the University of Brighton in 2016. The aim of the sustainability assessment was to understand the economic, environmental, and social benefits/disbenefits of transforming the former landfill into a public open space, managed long term (Li, Bardos, & Cundy, 2017), using SURF-UK qualitative sustainability assessment guidance (www.claire.co.uk/surfuk) enhanced with the HOMBRE idea of conceptual site models of sustainability (Bardos et al., 2016). Climate change-related considerations were a significant part of the sustainability assessment, including emissions of carbon to atmosphere versus sequestration; and economic factors such as the project’s future resilience. Unsurprisingly, the reuse of the capped landfill as a public park showed substantive sustainability improvement.

Anticipate Adsorb Reshape (“A2R”), a United Nations (2017) Climate Resilience Initiative can support sustainable, resilient cleanup and reuse of hazardous sites. A2R focuses on the capacity to reshape development pathways by: (1) transforming economies to reduce risks and root causes of vulnerabilities, and (2) supporting the sound management of physical infrastructure and ecosystems to foster climate resilience.

Complementing A2R is the World Bank vision of contaminated sites as “engines for economic development, sources of sustainable energy, food security & efficiency—all while assuring public health and environmental protection” (World Bank, 2009, p. 1).

7 | CONCLUSIONS

SURF’s recommendations can guide owners and project managers in developing a site resiliency strategy. By following a systematic, holistic approach with the site end use in mind, and by meeting priority social and economic needs, climate-resilient sustainable remediation and redevelopment can reduce public health risks and create long-term value for communities. SURF plans to partner with the private and public sector to support pilot studies and conduct national and international capacity-building.

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REFERENCES


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